

1 **Survival and Growth Factors Affecting Community-Planted Urban Street Trees**

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3 **ABSTRACT**

4 Urban street trees face adverse growing conditions: compacted soils, extreme heat, lack
5 of nutrients, drought, car damage and vandalism. Limited funding, however, is cited by urban
6 tree-planting organizations as their major obstacle. To maximize budgets, many organizations
7 along the eastern United States have planted bare root trees as a less expensive alternative to
8 balled-and-burlapped (B&B) trees. Existing research indicates equivalent survival rates between
9 bare root and B&B trees; but no research has examined this in community group-planted urban
10 street trees. Bare root trees are additionally advantageous in community-based plantings because
11 they are much lighter and easier for volunteers to handle. This study evaluated the influence of
12 stock and other site factors on street tree survival and growth measures (diameter at breast
13 height, percent canopy cover, and percent live crown), while controlling for species and age. Site
14 factors included street traffic intensity, site type (curbside, park, yard, or commercial corridor),
15 wound presence, and sidewalk pit cut dimensions. 1159 trees (representing ten species) planted
16 by Philadelphia community groups under the guidance of the Pennsylvania Horticultural Society
17 from 2006-2009 were sampled. Overall, trees showed a high survival rate of 95%, with no
18 significant difference between B&B and bare root trees. Species with the highest survival rates
19 were *Prunus virginiana* (chokecherry), *Platanus x acerifolia* (London plane tree), and *Acer*
20 *ginnala* (Amur maple). Heavily trafficked streets exhibited lower survival, percent canopy cover
21 and percent live crown. Larger growth measures were expected and found in B&B trees, as they
22 have historically been planted larger than their bare root counterparts. Findings support planting
23 larger trees (such as B&B and/or larger bare root trees) along commercial corridors. Species in
24 the Rosaceae family (*Amelanchier spp.*, *Malus spp.*, and *Prunus virginiana*) exhibited lower
25 percents canopy cover. Wound presence and pit cut size were not major factors affecting the 1-5
26 year old street trees sampled in this study. The major management implication of these findings
27 is that bare root trees are a viable alternative to B&B trees in community-based urban forestry
28 initiatives. Tree-planting campaigns with similar climactic conditions to Philadelphia can use this
29 study to inform selection of stock and species.

30

31 **KEYWORDS**

32 Urban forestry, harvesting method, bare root, balled-and-burlapped, B&B, Philadelphia, PHS,
33 TreeVitalize, transplanting, street traffic, wound, pit cut, *Acer ginnala*, *Acer rubrum*,
34 *Amelanchier*, *Cercis canadensis*, *Cladrastis kentukea*, *Gleditsia triacanthos*, *Malus*,
35 *Platanus x acerifolia*, *Prunus virginiana*, *Syringa reticulata*, Amur maple, red maple,
36 serviceberry, redbud, yellowwood, honey locust, crabapple, London plane tree, Japanese
37 tree lilac.

38

39 **INTRODUCTION**

40 Urban forests have been recognized for their role in improving the standard of living for
41 city residents across the country (Dwyer et al. 1992; Bolund & Hunhammer 1999; Sather et al.
42 2004). Street trees offer a wide array of services including improved air and water quality
43 (Beckett et al. 2000; Nowak et al. 2007), property value (McPherson et al. 1997), human health
44 (Coder 1996), energy conservation (Nowak 1995), wildlife habitat (Coder 1996), weather
45 buffering and urban heat island amelioration (McPherson 1994), storm water catchment
46 (McPherson et al. 1997), sense of community (Coder 1996), economic revitalization (Wolf
47 2003), and crime reduction (Kuo and Sullivan 2001). Crown fullness and size is often positively
48 correlated with these benefits; so that the larger, fuller trees produce greater effects (McPherson
49 et al. 1999). While urban reforestation campaigns have gained popularity in recent years, urban
50 street tree counts continue to decrease, and funding availability remains the greatest challenge
51 facing tree planting efforts (Kielbaso 1990; Hauer and Johnson 2008). Government agencies,
52 contractors, and non-governmental organizations across the country have thus been exploring
53 alternative means of maximizing the number of trees they can plant.

54 Bare root trees are seen as a less expensive, more easily transported alternative to balled-
55 and-burlapped (B&B) trees (Buckstrup & Bassuk 2003; Sather et al. 2004), and have been
56 planted in many cities in the eastern U.S.. The bare root method of transplanting trees has a long
57 standing history in nurseries, dating back to the industry's origin in the United States in the
58 eighteenth century (Davidson et al. 1999). Bare root trees are grown similarly to B&B trees but
59 are transplanted in a way such that the soil the trees are grown in can be shaken away, leaving
60 the roots exposed (Sather et al. 2004). In order to prevent the roots from desiccating, they are
61 recommended to be dipped into a hydrogel polymer slurry and wrapped in a clear plastic bag
62 (Buckstrup & Bassuk 2003; Harris et al. 2004). The roots are sensitive to changes in temperature,
63 moisture, and planting conditions, and should therefore be transplanted while dormant during
64 spring or fall, and within a week of shipping from the nursery (Sather et al. 2004). Bare root trees
65 have many advantages as they are a fraction of the weight and cost of B&B trees, can be shipped
66 more efficiently, and root pruned for visible defects prior to planting (Buckstrup and Bassuk
67 2003; Flott et al. 2008). In sum, bare root plantings allow more volunteers to plant more trees
68 within the same constraints of community group capacity and funding availability.

69 Initial research directly comparing bare root and B&B trees by Cool (1976) found higher
70 mortality in bare root trees than in B&B trees. A follow-up study by Vanstone and Ronald
71 (1981) found that if transplanted correctly, no difference in mortality was evident between stocks
72 by the second growing season. B&B trees did however score higher in growth indices (shoot
73 growth and leaf size). Buckstrup and Bassuk (2000) conducted a similar study directly
74 comparing the mortality and growth rates of bare root trees to those of B&B trees over two
75 growing seasons. Their mortality findings substantiated those put forth by Vanstone and Ronald
76 (1981), but their data on growth indices indicated no differences across stock. Most recently,

77 Anella et al. (2008) further corroborated Buckstrup & Bassuk's (2000) findings in the more
78 drought-inclined environment of Oklahoma. All of these studies emphasized the importance of
79 sampling across species and growing seasons. Despite these findings, popular belief still holds
80 that B&B trees consistently have higher survival rates than their bare root counterparts (Sather et
81 al. 2004). The body of research on urban bare root trees currently lacks any studies directly
82 comparing bare root and B&B trees planted by volunteers.

83 The main objective of this study is to highlight the role stock (bare root vs. B&B) plays in
84 the survival and growth of community-planted street trees. I prioritized controlling for species
85 variability and sampling across multiple growth seasons. Secondary factors under consideration
86 are street traffic intensity, site type (curbside, yard, park, and commercial corridor), wound
87 presence, and dimensions of sidewalk pit cuts. Street traffic (both pedestrian and automobile) can
88 in theory impact tree survival and growth because they are at a higher risk of damage from cars
89 and vandalism. Site type can impact a tree's access to water, as well as its exposure to traffic-
90 related risks. Wounding opens a tree up to infection and can therefore impact a tree's survival
91 and growth. Sidewalk pit cut size in theory acts as a proxy for access to rainwater as well as
92 other constraints on root growth. The role of these factors has not been quantified in the existing
93 body of published research on community urban forestry.

94 I hypothesize that bare root and B&B trees will have equivalent survival rates. I predict
95 that across stock, higher rates of mortality will be positively correlated with smaller pit cut size,
96 wound presence and higher street traffic. Because bare root trees are often specified to be planted
97 at smaller caliper size than B&B trees, I hypothesize that growth measures will be
98 correspondingly larger in B&B trees. Higher growth measures are also expected along less

99 trafficked streets, in yard trees, in trees without wounding, and in larger pit cuts. Species-specific
100 variability in DBH and crown fullness is also expected based on tree habit and form.

101

102 **METHODS**

103 **Site Selection and Sampling Design**

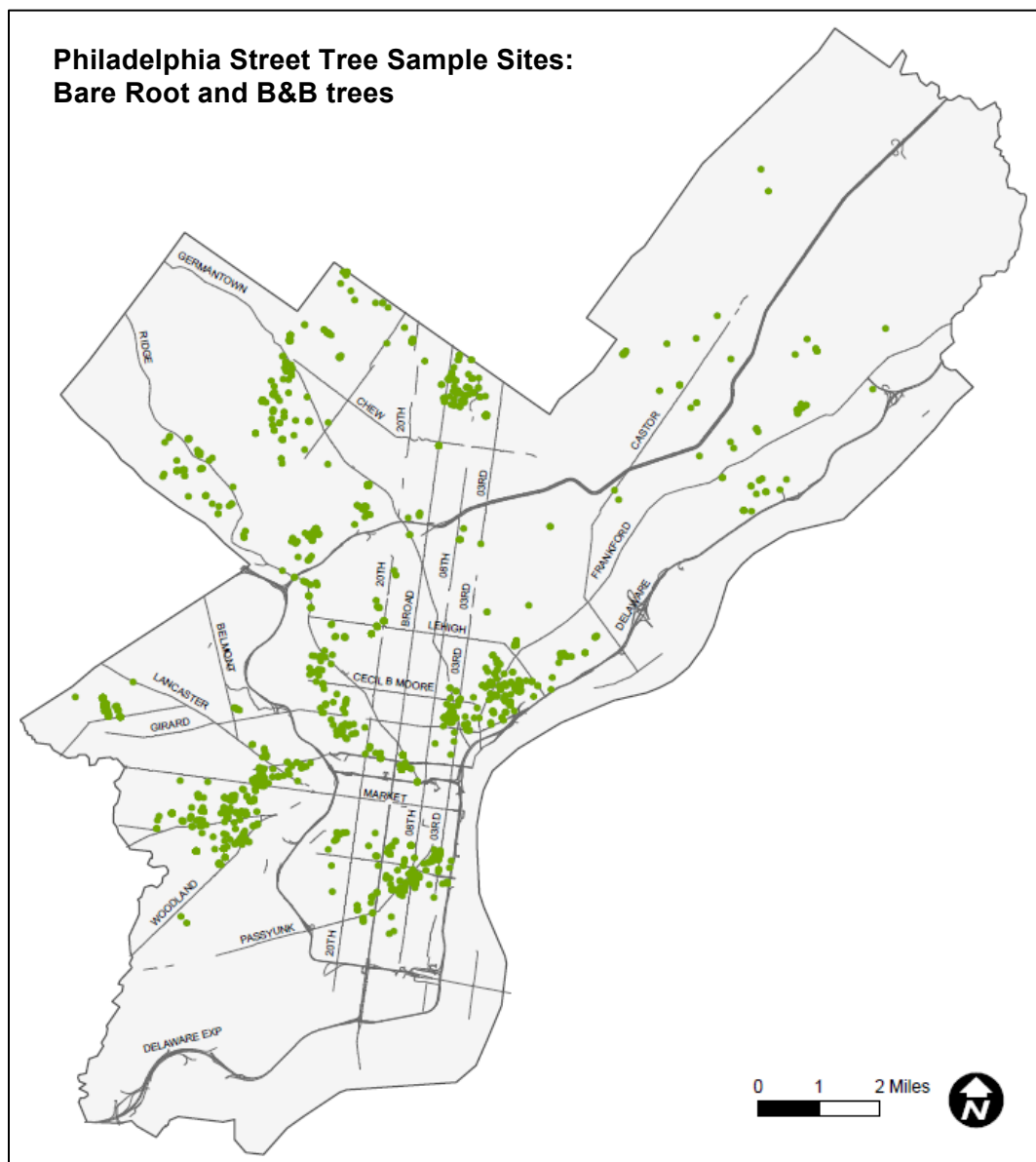
104 This study was conducted on trees planted in Philadelphia through the southeast
105 Pennsylvania TreeVitalize campaign coordinated by the Pennsylvania Horticultural Society
106 (PHS). Philadelphia is located at 39° 57' 8" N / 75° 9' 51" W along the mid-Atlantic border of
107 the United States. The city covers 326.14 km² and is situated at an elevation of 11.89 m above
108 mean sea level. It is home to 1.5 million people and 2.1 million trees. The city has an average
109 canopy cover of 15.7% (ranging from 1.8% to 38.3% by neighborhood) (American Forests
110 2003), and more than half of its trees have diameter at breast height (DBH) sizes of less than
111 15.25 cm (6 in) (Nowak et al. 2007). Every year hundreds of trees are planted by the city
112 government through the Department of Parks and Recreation (through a division formerly called
113 the Fairmount Park Commission). Many hundreds, and in recent years thousands, more are
114 planted by PHS through the TreeVitalize campaign.

115 TreeVitalize is a public-private partnership launched by the PA Department of
116 Conservation and Natural Resources in 2004. In less than five years the program reached its
117 initial goal of planting over 20,000 trees in and around Philadelphia through community
118 members. The program continues to grow in southeast PA and has now been launched in all
119 other metropolitan areas across the state. Tree-planting volunteers are led by community group
120 leaders trained through the 9-hour Tree Tender® course developed by PHS in collaboration with
121 Penn State Cooperative Extension. The training program covers tree planting, identification,

122 benefits, and maintenance. The fact that TreeVitalize represents both public and private
123 organizations, as well as community groups makes it an ideal urban forestry program for study.

124 PHS has maintained records of every tree planted since 2004, including detailing species,
125 planting address, stock (bare root vs. B&B), and the community group that planted it. An
126 analysis of these records showed the most commonly planted species in both bare root and B&B
127 stocks have been Amur maple (*Acer ginnala* Maxim., Sapindaceae), red maple (*Acer rubrum* L.,
128 Sapindaceae), serviceberry (*Amelanchier* Medik., Rosaceae), redbud (*Cercis canadensis* L.,
129 Rosaceae), yellowwood (*Cladrastis kentukea* (Dum. Cours.) Rudd, Fabaceae), honey locust
130 (*Gleditsia triacanthos* L., Fabaceae), crabapple (*Malus* Mill., Rosaceae), London plane tree
131 (*Platanus x acerifolia* Willd., Platanaceae), chokecherry (*Prunus virginiana* L., Rosaceae), and
132 Japanese tree lilac (*Syringa reticulata* (Blume) H.Hara, Oleaceae). Records for these ten species
133 were sorted by stock and year planted, and then randomized. Up to 30 sites for each stock, of
134 each species, from each planting year, were randomly selected (some groupings were limited to
135 fewer than 30 sites). The sampling design was fully crossed, and blocked by planting year.

136 Because some neighborhoods are more involved in TreeVitalize plantings than others, sites were
137 not distributed evenly across the city. A total of 1411 sites were selected through this process and
138 mapped using ArcGIS (Figure 1). Of the original 1411 trees sites, 644 (45%) were B&B trees,
139 and 767 (55%) were bare root trees.



140

141 **Figure 1:** Based on records from the Pennsylvania Horticultural Society, 1411 bare root
 142 and balled-and-burlapped trees were randomly sampled for mortality and growth in metropolitan
 143 Philadelphia. Sites were chosen across ten species and four age classes from 2006-2009.

144

145 **Data Collection**

146 Site inspections were conducted between mid-June and early August to ensure that all
 147 trees were fully leafed-out upon time of inspection. At each site, address and species planted

148 were verified or revised. Three growth/vigor measures were then taken: diameter at breast height
149 (DBH), percent canopy cover, and percent live crown. DBH was measured in quarter-inch
150 increments using calipers. Four measurements of percent canopy cover were taken using a
151 densiometer. These measurements were taken at curbside, left, right, and sidewalk-facing
152 directional points; and averaged to capture variability. In order to calculate percent live crown,
153 trunk height and total height were measured in half-foot and foot increments, respectively.
154 Measurements followed Colorado State Forest Service guidelines (Schomaker 2004). Percent
155 live crown was later calculated by dividing live crown height (total height – trunk height) by
156 total height and multiplying by 100. Street traffic intensity (residential low traffic, residential
157 high traffic, or commercial) was based on visual assessment of site land use and traffic intensity
158 (both vehicular and pedestrian). Site type (commercial corridor, curbside, yard, park), tree trunk
159 wound presence, and sidewalk pit cut dimensions were also recorded.

160 **Statistical Analysis**

161 Statistical analysis was done using R statistical software. DBH data was normalized using
162 a log transformation, while percent canopy cover and percent live crown were normalized using
163 an $\arcsin(\sqrt{\cdot})$ transformation. Logistic regression was done to analyze mortality data (Packer and
164 Clay 2000), and multiple linear regressions in conjunction with regression trees were used to
165 analyze growth measure data (DBH, percent canopy cover, percent live crown) (Gregg et al.
166 2003). Regression trees were used as a visual aid in determining significant interaction effects
167 (De'ath and Fabricius 2000). Explanatory variables included stock, species, age, site traffic, site
168 type, wound presence, and pit cut size. I used a forward selection procedure to retain only those
169 factors that were significant in my model (Peña-Claros et al. 2008, De'ath and Fabricius 2000).
170 A two-way ANOVA with an error term for years planted (age) was also done to test for an

171 interaction between species and stock while accounting for the blocked sampling design (Peña-
172 Claros et al. 2008).

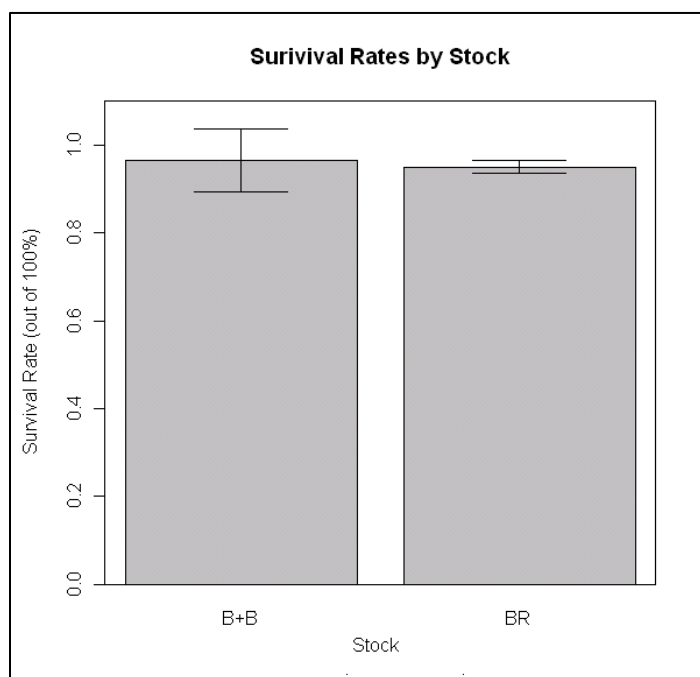
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174 **RESULTS**

175 A majority of the 1411 tree sites selected were found and inspected (1159, 82.6%). Those
176 not found were either, a) never planted, or b) planted, died and removed. 89% of the B&B sites
177 and 77% of bare root trees were found. While this suggests a potentially higher rate of mortality
178 in bare root trees, in reality this discrepancy is due to a higher rate of clerical mistakes in early
179 bare root planting years. Sites not found were not included in statistical analysis.

180 **Mortality**

181 Both bare root and B&B stocks exhibited very high survival rates: 96% of B&B trees,
182 and 95% of bare root trees after an average of 2.62 years since planting. The results of logistic
183 regression analysis indicated no significant difference in survival rates between bare root and
184 B&B trees (Figure 2).

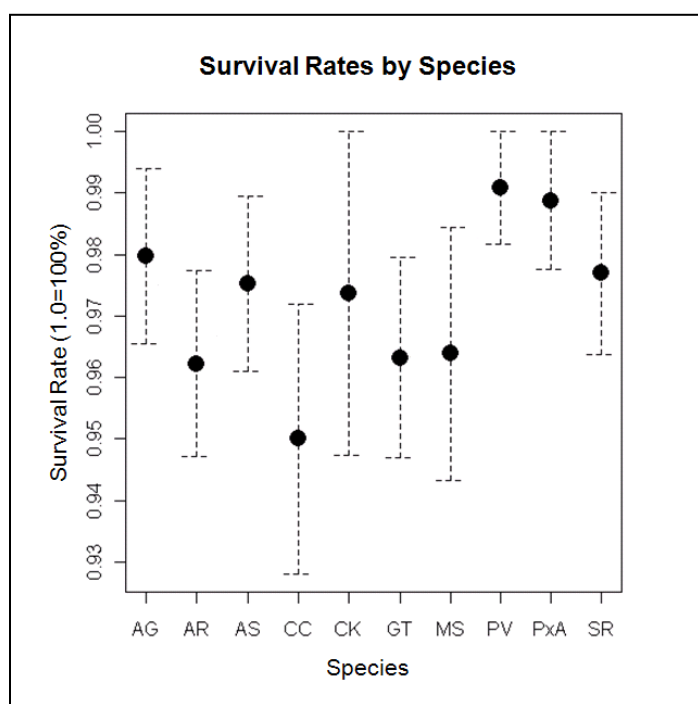


185

186 **Figure 2** High rates of survival were found both stocks; with no significant difference in survival
187 rates between B&B and bare root trees.

188
189 Years since planting (“age”) were also not a significant predictor of mortality ($p=0.921$).

190 Species with the highest survival rates were *P. virginiana* and *Platanus x acerifolia*, with *A.*
191 *ginnala*, *S. reticulata*, *Amelanchier spp.*, and *C. kentukea* close behind; *C. canadensis* had the
192 lowest survival rate (Figure 3).



193
194 **Figure 3** Species sampled were *Acer ginnala* (AG), *Acer rubrum* (AR), *Amelanchier spp.* (AS),
195 *Cercis canadensis* (CC), *Cladrastis kentukea* (CK), *Gleditsia triacanthos* (GT), *Malus spp.*
196 (MS), *Prunus virginiana* (PV), *Platanus x acerifolia* (PxA), and *Syringa reticulata* (SR).
197 *Prunus virginiana* (chokecherry) and *Platanus x acerifolia* (London plane tree) had the highest
198 survival, and *Cercis canadensis* (redbud) had the lowest.

200 Street traffic intensity was the only other significant factor to affect survival rates. Lower
 201 survival rates were observed along heavily trafficked commercial corridors (p=0.0056) (Table 1).
 202 **Table 1** –Higher mortality was found along commercial corridors (the highest traffic intensity).
 203 *Prunus virginiana* (chokecherry) exhibited the highest rate of survival, while *Cercis Canadensis*
 204 (redbud) exhibited the lowest; although all species showed very high rates of survival (>93%).
 205 Significant codes: 0 ‘***’, 0.001 ‘**’, 0.01 ‘*’.

Survival Coefficients (Equivalent R-squared = 0.124, AIC = 379.53)		
	Estimate	P-value
Commercial Traffic	-0.991	0.0056**
<i>Acer ginnala</i>	2.537	0.0009***
<i>Acer rubrum</i>	1.812	0.0002***
<i>Amelanchier spp.</i>	2.254	0.0004***
<i>Cercis canadensis</i>	1.481	0.0051**
<i>Cladrastis kentukea</i>	2.404	0.0225*
<i>Gleditsia triacanthos</i>	1.978	0.0002***
<i>Malus spp.</i>	1.859	0.0039**
<i>Prunus virginiana</i>	3.296	0.0015**
<i>Platanus x acerifolia</i>	3.065	0.0032**
<i>Syringa reticulata</i>	2.424	0.0002***

206 **Growth**
 207

208 Growth was evaluated using three measures: diameter at breast height (DBH), percent
 209 canopy cover, and percent live crown. Due to discrepancies between recorded and actual species
 210 planted on sites, only *Acer rubrum* (red maple), *Amelanchier spp.* (serviceberry), and *Gleditsia*
 211 *triacanthos* (honey locust) were found in sufficient quantities to be included in two-way
 212 ANOVA analysis. Results indicated that DBH was significantly affected by stock (p_[1,15]=0.020)
 213 and species (p_[2,15]=0.029). Percent canopy cover was significantly affected by species
 214 (p_[2,15]<0.0001). Tukey tests indicated that with regard to DBH, *Amelanchier spp.* vs. *A. rubrum*
 215 (p<0.0001) as well as *G. triacanthos* vs. *Amelanchier* (p<0.0001) were significantly different;

216 while *A. rubrum* and *G. triacanthos* were not ($p=0.913$). All three species pairings had
 217 significantly different percent canopy covers (all $p<0.0001$).

218 According to multiple regression analyses, older and B&B trees were found to have
 219 larger measures for all three growth indices ($p<0.0001$). This was expected as trees obviously
 220 grow larger with age, and because B&B trees are initially planted larger than bare root stock.
 221 Growth indices also varied significantly with species. *Platanus x acerifolia* ($p<0.0001$) had
 222 significantly higher DBH values, while *Amelanchier spp.* ($p<0.0001$), *Malus spp.* ($p=0.0005$),
 223 and *Syringa reticulata* ($p=0.0003$) had lower DBH measures. Pit cut size was surprisingly
 224 negatively correlated with DBH in *Platanus x acerifolia* trees ($p=0.0070$). Age, stock, species,
 225 and pit cut size accounted for 51.9% of the variability in DBH measures (Table 2).

226 **Table 2** – Age, stock, species, and pit cut size were significant explanatory variables affecting
 227 DBH. Higher DBH measures were recorded in older, B&B, and *Platanus x acerifolia* (London
 228 plane tree) trees. Lower measures were recorded in trees that were younger, bare root,
 229 *Amelanchier* (serviceberry), *Malus spp.* (crabapple), and *Syringa reticulata* (Japanese tree lilac).
 230 Significant codes: 0 ‘***’, 0.001 ‘**’, 0.01 ‘*’.

DBH Regression Coefficients (Adjusted R-squared = 0.519)		
	Estimate	P-value
Age	0.1856	<0.0001***
Stock	-0.1418	<0.0001***
<i>Amelanchier spp.</i>	-0.2718	<0.0001***
Age:<i>Malus spp.</i>	-0.0339	0.0005***
Age:<i>Platanus x acerifolia</i>	0.1636	<0.0001***
Age:<i>Syringa reticulata</i>	-0.0429	0.0003***
Age:<i>Platanus x acerifolia</i>:Pit Cut Size	-0.0029	0.0070**

231
 232 Percent canopy cover was significantly correlated with species, street traffic intensity and
 233 wound presence. *A. ginnala* ($p<0.0001$), *Amelanchier spp.* ($p<0.0001$), *G. triacanthos*

234 (p<0.0001), *Malus spp.* (p<0.0001), and *P. virginiana* (p<0.0001) had significantly lower percent
 235 canopy cover measures than other species. *G. triacanthos* trees with bole wounds also had lower
 236 percent canopy cover (p=0.0416). Trees on low traffic residential streets had higher percents
 237 canopy cover (p=0.0011) than those on high traffic residential streets and commercial corridors.
 238 These factors accounted for 40.6% of variance in percent canopy cover data (Table 3).
 239 **Table 3**– Higher percent canopy cover measures were recorded in older, B&B, and along low
 240 traffic residential streets. Lower measures were recorded in trees that were younger, bare root,
 241 *Acer ginnala* (Amur maple), *Amelanchier* (serviceberry), *Gleditsia triacanthos* (honey locust),
 242 *Malus spp.* (crabapple), and *Prunus virginiana* (crabapple). Significant codes: 0 ‘***’, 0.001
 243 ‘**’, 0.01 ‘*’.

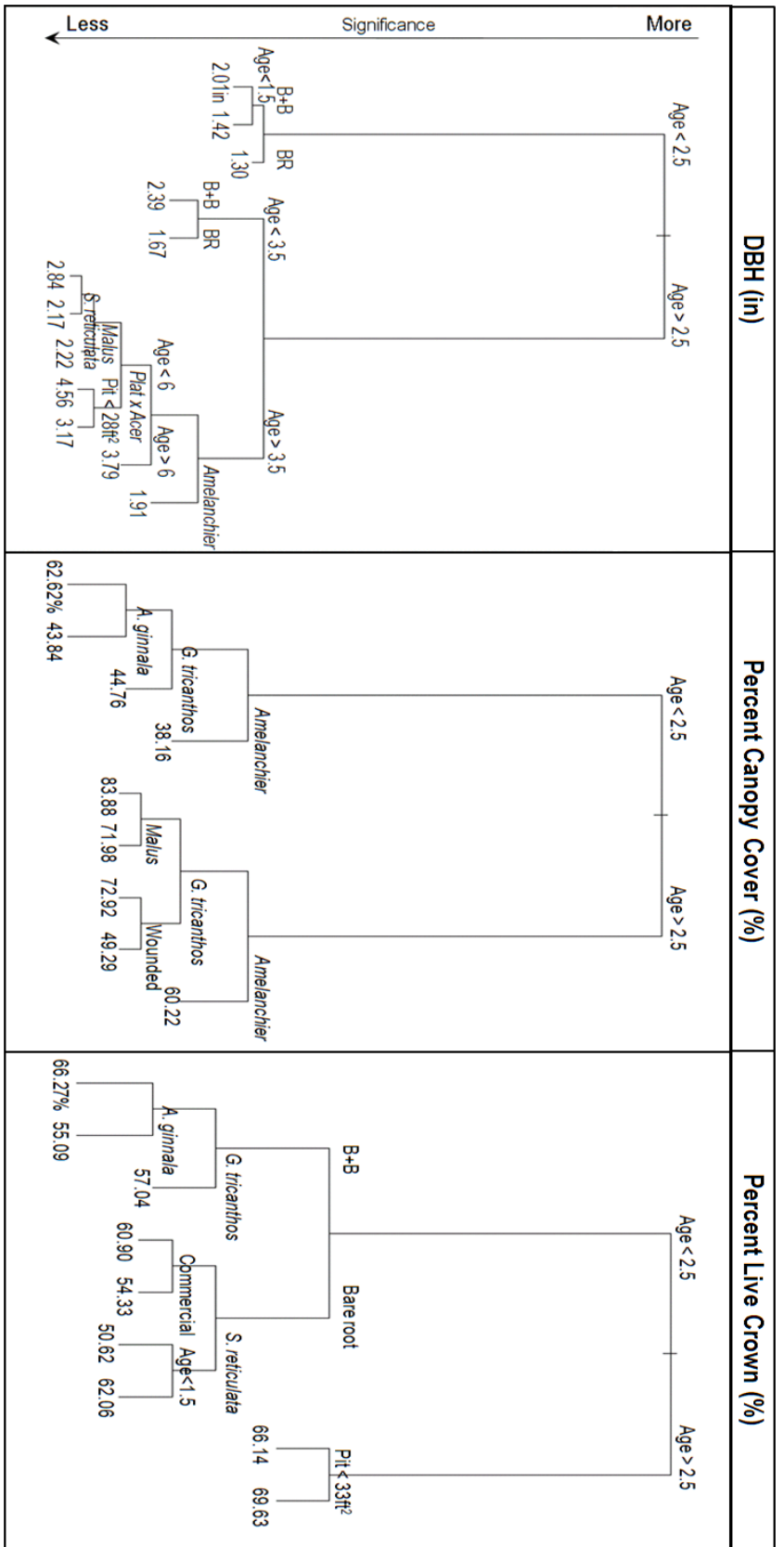
Percent Canopy Cover Coefficients (Adjusted R-squared=0.406)		
	Estimate (x10⁻⁵)	P-value
Age	0.916	<0.0001***
Stock	-0.635	<0.0001***
<i>Acer ginnala</i>	-1.288	<0.0001***
<i>Amelanchier spp.</i>	-3.185	<0.0001***
<i>Gleditsia triacanthos</i>	-2.206	<0.0001***
<i>Malus spp.</i>	-1.844	<0.0001***
<i>Prunus virginiana</i>	-1.063	<0.0001***
Low Residential traffic	0.446	0.0011**
<i>G. triacanthos</i>:Wound Presence	-1.019	0.0416*

244
 245 Higher percent live crown was found in curbside (p=0.0271), yard (p=0.0017), *P.*
 246 *virginiana* (p=0.0026) and *Platanus x acerifolia* (p=0.0397) trees. Lower percents live crown
 247 were observed in *A. ginnala* (p=0.018), *C. kentukea* (p<0.0001), *G. triacanthos* (p=0.0002), and
 248 *S. reticulata* (p<0.0001). Younger bare root trees (except *S. reticulata*) along commercial
 249 corridors had lower percents live crown (p=0.0003), though this trend diminished with age.
 250 Older trees in smaller sidewalk pit cuts (p=0.0049), and wounded *G. triacanthos* trees with

251 (p=0.022) also had lower percents live crown. Age, stock, species, pit cut size, wound presence,
 252 site type and street traffic accounted for 25% of variance in percent live crown (Table 4).
 253 **Table 4**—Higher measures were recorded in older, B&B, street, yard, *Platanus x acerifolia*
 254 (London plane tree) and *Prunus virginiana* (chokecherry) trees. Lower measures were recorded
 255 in trees that were younger, bare root, in smaller pits, with wounds, and along commercial
 256 corridors. Significant codes: 0 ‘***’, 0.001 ‘**’, 0.01 ‘*’.

Percent Live Crown Coefficients (Adjusted R-squared=0.2505)		
	Estimate (x10⁻⁶)	P-value
Age	2.628	<0.0001***
Stock	-2.331	<0.0001***
Pit Cut Size	0.089	0.0049**
Wound Presence	-1.949	0.022*
<i>Acer ginnala</i>	-4.098	0.018*
<i>Cladrastis kentukea</i>	-4.998	<0.0001***
<i>Gleditsia triacanthos</i>	-3.722	0.0002***
<i>Prunus virginiana</i>	3.253	0.0026**
<i>Platanus x acerifolia</i>	2.467	0.0397*
<i>Syringa reticulata</i>	-5.234	<0.0001***
Street/Curbside Site Type	1.734	0.0271*
Yard Site Type	4.218	0.0017**
Stock:Commercial Traffic	-6.235	0.0003***

257
 258 The visual interactions of significant factors affecting all growth measures are well
 259 illustrated through regression trees (Figure 4). Regression trees display significance from the top
 260 down, with longer branches indicating higher significance.



263 **Figure 4** – Regression trees illustrate significant interactions between factors affecting growth
264 measures. They can be read from the top down; longer branches indicate more significant trends.
265 For example, regarding percent live crown, trees under the age of 2.5 years, that are bare root, a
266 species other than *Syringa reticulata* (Japanese tree lilac), and located along a commercial
267 corridor has average percent live crown of 54.33%. This is significantly lower than trees that
268 meet the same criteria but are not located along commercial corridors (60.90%).

269

270 **DISCUSSION**

271 The main objective of this study was to investigate the influence of stock (B&B vs. bare
272 root) on survival and growth of urban street trees planted by community groups. Both bare root
273 and B&B displayed very high survival rates, with no significant difference across stock. This
274 substantiates findings by Vanstone and Ronald (1981), Buckstrup and Bassuk (2000), and Anella
275 (2008) within the context of community-based plantings.

276 With regard to growth measures, B&B trees were found consistently to have higher
277 DBH, percent canopy cover, and percent live crown values. This was expected because B&B
278 trees are regularly planted larger than their bare root counterparts. Because this study was based
279 on previously planted trees, size at which they were planted was not standardized across stocks.
280 This study cannot provide insight into the comparative rates of growth; however, it does
281 highlight that B&B trees currently have larger DBH measures and fuller crowns on average.
282 Significance of stock on DBH dissipates with age; such that DBH of older trees becomes more
283 correlated with species (Figure 4). This supports Buckstrup and Bassuk's (2000) finding that it
284 may take multiple growing seasons for size differences between stocks to level out.

285 Findings confirmed variability across species. For example, *Amelanchier spp.* was
286 consistently smaller than the other species, while *Platanus x acerifolia* was much larger. Species-
287 specific variability is seen as a reflection of variance in habit and form rather than performance.
288 All species sampled from the Rosaceae family (*Amelanchier spp.*, *Malus spp.*, and *Prunus*
289 *virginiana*) had lower percent canopy cover measures. This may have implications for species
290 selection as percent canopy cover is often used by organizations in setting tree cover goals.

291 Explanatory variables accounted for the most variance in DBH (51.9%), followed by
292 percent canopy cover (40.6%), and percent live crown (25%). Street traffic intensity was the
293 most significant factor in these models after age, stock, and species. Given that wound presence
294 was significant in percent live crown, it may be that with greater street traffic more passersby are
295 breaking off branches, or creating other sources of stress. Under stress, the tree may not be able
296 to allocate as many resources towards crown fullness.

297 While only a quarter of the variance in percent live crown was explained through the
298 statistical model used, the trends it exhibits are nonetheless insightful. High percent live crown
299 was seen in yard trees – higher than in park, curbside, or corridor trees. Yard trees may have
300 greater resource availability, and ability to allocate carbohydrates towards secondary growth.
301 Species-specific form was also underlined by findings, with large live crowns in *Platanus x*
302 *acerifolia* and *Prunus virginiana*.

303 Pit cut size had a much smaller influence on mortality and growth measures than was
304 hypothesized. Pit cut size was only positively correlated with percent live crown in trees 2.5
305 years after planting. Further investigation into pit cut dimension influence on street trees with
306 time would shed light on if this trend becomes more pronounced in more mature trees when root
307 growth becomes more constricted. It should be noted that this study was conducted on ten

308 commonly planted street tree species, so results for species less well-suited to the urban
309 environment may vary in terms of survival and growth rates. Research on less commonly-planted
310 urban tree species, as well as the significance of soil type should also be pursued in the future.

311

312 **MANAGEMENT IMPLICATIONS**

313 Tree-planting campaigns in greater Philadelphia, and other areas with similar planting
314 conditions, can use this study in decision-making when selecting street tree stock and species
315 (especially along commercial corridors). This study's primary management implication is that
316 bare root tree plantings can be carried out without concern for higher mortality. This has huge
317 implications not only for organizations like PHS hoping to maximize tree-planting budgets and
318 volunteer involvement; but also for nurseries in the area. This study ameliorates concerns held by
319 local nurseries that bare root trees do not survive as well as B&B trees in the urban environment.

320 Higher DBH, percent canopy cover and percent live crown are useful to think about in
321 terms of maximizing the benefits garnered from planting street trees. For example, air quality
322 filtration, and storm water catchment are notably improved in larger trees, with fuller crowns
323 (McPherson et al. 1999). Findings from this study therefore uphold *Platanus x acerifolia* as a
324 highly beneficial street tree. This study indicates that larger trees performed well along heavily
325 trafficked streets, and because high traffic intensity affects survival, a concerted effort should be
326 made to either plant B&B or larger bare root trees along commercial corridors.

327 Lastly, this study emphasizes the importance of exploring urban forestry research in
328 street trees planted by community groups. This research model can provide insight into trends for
329 urban tree-planting organizations elsewhere. Accurate, comprehensive record-keeping is highly

330 encouraged in order to support future research. A follow-up study on the same trees measured for
331 this study would generate valuable information about growth and survival over time.

332

333 **REFERENCES**

334 American Forests. 2003. Urban ecosystem analysis: Delaware valley region. Special publication
335 of American Forests and the USDA Forest Service. Washington, DC. 12pp.

336 Anella, L., Hennessey T.C., and E.M. Lorenzi. 2008. Growth of balled-and-burlapped vs. bare-
337 root trees in Oklahoma, US. *Journal of Arboriculture & Urban Forestry* 34(3):200-203.

338 Beckett, K.P., P. Freer-Smith, and G. Taylor. 2000. Effective tree species for local air-quality
339 management. *Journal of Arboriculture* 26:12–19.

340 Bolund, P. and Hunhammar, S. 1999. Ecosystem services in urban areas. *Ecological Economics*
341 29:293-301.

342 Buckstrup, M.J. and Bassuk, N.L. 2000. Transplanting success of balled-and-burlapped versus
343 bare-root trees in the urban landscape. *Journal of Arboriculture* 26(6):298-307.

344 ----- . 2003. *Creating the Urban Forest: The Bare Root Method*. Cornell University's Urban
345 Horticulture Institute, Ithaca, NY. 18pp.

346 Burden, D. 2008, *22 Benefits of Urban Street Trees*, Glattig Jackson and Walkable
347 Communities University of Montana, Missoula, MT.

348 Cool, R.A. 1975. Tree Spade vs. Bare Root Tree Planting. *Journal of Arboriculture* 2(5):92-95.

349 Coder, K.D. 1996. *Identified Benefits of Community Trees and Forests*. University of Georgia
350 School of Forest Resources. Athens, GA. 7pp.

351 Davidson, H., Mecklenberg, R. and Peterson, C. 1999 (4th edition). *Nursery Management*.
352 Prentice Hall, Englewood Cliffs, NJ. 530pp.

353 De'ath, G., and K.E. Fabricius. 2000. Classification and regression trees: A powerful yet simple
354 technique for ecological data analysis. *Ecology* 81(11):3178-3192.

355 Dwyer, J.F., McPherson, E.G., Schroeder, H.W. and Rowntree, R.A. 1992. Assessing the
356 benefits and costs of the urban forest. *Journal of Arboriculture* 18(5):227-234.

357 Ferrini, F., Nicese, F.P., Mancuso, S. and Giuntoli, A. 2000. Effect of nursery production method
358 and planting techniques on tree establishment in urban sites: preliminary results. *Journal*
359 *of Arboriculture* 26(5):281-284.

360 Flott, J., Appleton, B. and Baker, R.B. 2008. Bare Rooting – A Planting and Transplanting
361 Technique. Presented at the 2008 International Society of Arboriculture annual
362 conference. 16pp. [http://www.indiana-arborist.org/documents/pdf/BareRooting-](http://www.indiana-arborist.org/documents/pdf/BareRooting-FlottAppletonBaker.pdf)
363 [FlottAppletonBaker.pdf](http://www.indiana-arborist.org/documents/pdf/BareRooting-FlottAppletonBaker.pdf). Accessed 5/10/2010.

364 Gregg, J. W., Jones, C. G. and T. E. Dawson. 2003. Urbanization effects on tree growth in the
365 vicinity of New York City. *Nature* 424:183– 187.

366 Harris, R.W., Clark, J.R. and Matheny, N.P. 2004 (4th edition). *Arboriculture: Integrated*
367 *Management of Landscape Trees, Shrubs, and Vines*. Prentice Hall. New York, NY.

368 Hauer, R.J. and Johnson, G.R. 2008. State urban and community forestry program funding,
369 technical assistance, and financial assistance within the 50 United States. *Journal of*
370 *Arboriculture & Urban Forestry* 34(5):280-289.

371 Kielbaso, J.J. 1990. Trends and Issues in City Forests. *Journal of Arboriculture* 16(3):69-76.

372 Kuo, F.E. and Sullivan, W.C. 2001. Environment and Crime in the Inner City: Does vegetation
373 reduce crime?. *Environment and Behavior* 33(3):343-367.

374 McPherson, E. G. 1994. Cooling urban heat islands with sustainable landscapes, pp. 151–71. In
375 (R. H. Platt, R. A. Rowntree, and P. C. Muick, Eds.). *The Ecological City, Preserving and*
376 *Restoring Urban Biodiversity*. University of Massachusetts Press, Boston, MA.

377 McPherson, E.G., Nowak, D., Heisler, G., Grimmond, S., Souch, C., Grant, R. and R. Rowntree.
378 1997. Quantifying urban forest structure, function, and value: the Chicago urban forest
379 climate project. *Urban Ecosystems* 1:49–61.

380 McPherson, E.G., Simpson, J.R., Peper, P.J., and Q. Xiao. 1999. Benefit-cost analysis of
381 Modesto’s municipal urban forest. *Journal of Arboriculture* 25(5):235–248.

382 Nowak, D.J.. 1995. Trees pollute? A “TREE” explains it all, pp. 28-30. In (C. Kollin and M.
383 Barratt, Eds.). *Proceedings from the 7th National Urban Forestry Conference, American*
384 *Forests*. Washington, DC.

385 Nowak, D.J., Hoehn R.E., Crane D.E., Stevens J.C., and Walton J.T. 2007. *Assessing Urban*
386 *Forest Effects and Values: Philadelphia Urban Forest*. USDA Forest Service, Newtown
387 Square, PA. 28pp.

388 Packer, A. and K. Clay. 2000. Soil pathogens and spatial patterns of seedling mortality in a
389 temperate tree. *Nature* 404:278–281

390 Peña-Claros, M., Fredericksen, T.S., Alarcón, A., Blate, G.M., Choque, U., Leñaño, C., and J. C.
391 Licona. 2008. Beyond reduced-impact logging: Silvicultural treatments to increase
392 growth rates of tropical trees. *Forest Ecology and Management* 256:1458–1467.

393 Sather, I., Macie, E., and Hartel, D.R. 2004. *Urban Forestry Manual – Benefits and Costs of the*
394 *Urban Forest*. USDA Forest Service, Southern Center for Urban Forest Research &
395 Information. Athens, GA. 377 pp.

396 Schomaker, M. 2004. Section 12: Crowns: Measurements and Sampling. Phase 3 Field Guide.
397 Colorado State Forest Service, Fort Collins, CO, USA.
398 Vanstone, D.E. and W.G. Ronald. 1981. Comparison of bareroot verses tree spade transplanting
399 of boulevard trees. *Journal of Arboriculture* 7(10):271-274.
400 Wolf, K. L., 2003. Public response to the urban forest in inner-city business districts, *Journal of*
401 *Arboriculture* 29(3):117–126.

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